

# Temperatures in Excess of Critical Thresholds Threaten Nestling Growth and Survival in a Rapidly-Warming Arid Savanna: A Study of Common Fiscals

Susan J. Cunningham\*, Rowan O. Martin, Carryn L. Hojem, Philip A. R. Hockey†

Percy FitzPatrick Institute, DST/NRF Centre of Excellence, University of Cape Town, Rondebosch, South Africa

## Abstract

Frequency, duration, and intensity of hot-weather events are all predicted to increase with climate warming. Despite this, mechanisms by which temperature increases affect individual fitness and drive population-level changes are poorly understood. We investigated the link between daily maximum air temperature ( $t_{\max}$ ) and breeding success of Kalahari common fiscals (*Lanius collaris*) in terms of the daily effect on nestling body-mass gain, and the cumulative effect on size and age of fledglings. High  $t_{\max}$  reduced mass gain of younger, but not older nestlings and average nestling-period  $t_{\max}$  did not affect fledgling size. Instead, the frequency with which  $t_{\max}$  exceeded critical thresholds ( $t_{\text{crits}}$ ) significantly reduced fledging body mass ( $t_{\text{crit}} = 33^{\circ}\text{C}$ ) and tarsus length ( $t_{\text{crit}} = 37^{\circ}\text{C}$ ), as well as delaying fledging ( $t_{\text{crit}} = 35^{\circ}\text{C}$ ). Nest failure risk was 4.2% per day therefore delays reduced fledging probability. Smaller size at fledging often correlates with reduced lifetime fitness and might also underlie documented adult body-size reductions in desert birds in relation to climate warming. Temperature thresholds above which organisms incur fitness costs are probably common, as physiological responses to temperature are non-linear. Understanding the shape of the relationship between temperature and fitness has implications for our ability to predict species' responses to climate change.

**Citation:** Cunningham SJ, Martin RO, Hojem CL, Hockey PAR (2013) Temperatures in Excess of Critical Thresholds Threaten Nestling Growth and Survival in a Rapidly-Warming Arid Savanna: A Study of Common Fiscals. PLoS ONE 8(9): e74613. doi:10.1371/journal.pone.0074613

**Editor:** Justin G. Boyles, Southern Illinois University, United States of America

**Received:** May 4, 2013; **Accepted:** August 5, 2013; **Published:** September 9, 2013

**Copyright:** © 2013 Cunningham et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

**Funding:** This study was funded by a grant from the Department of Science and Technology (<http://www.dst.gov.za/>) through the National Research Foundation (<http://www.nrf.ac.za/>) of South Africa to the Centre of Excellence (CoE) in Birds as Keys to Biodiversity Conservation at the Percy FitzPatrick Institute of African Ornithology (<http://www.fitzpatrick.uct.ac.za/docs/coe.html>). Additional funding from the Tswalu Foundation (<http://www.tswalu.com/blog/the-tswalu-foundation/>) covered accommodation costs at Tswalu Kalahari Reserve. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

**Competing Interests:** The authors have declared that no competing interests exist.

\* E-mail: [susie.j.c@gmail.com](mailto:susie.j.c@gmail.com)

† Deceased.

## Introduction

Climate change is causing range shifts and local extinctions of species worldwide [1–3], particularly at the ‘warm edges’ of their ranges [4]. Accurately predicting and understanding such changes relies upon knowledge of the mechanistic links between climate and key biological processes [5–7]. Empirical data on how organisms respond to climatic variation and the implications for fitness are therefore of great importance [8].

Animals living in hot, arid environments routinely face harsh climatic conditions, including environmental temperatures with the potential to induce lethal hyperthermia (see [9–11] for examples of mass mortalities of bats and birds during heat-waves). Endotherms respond to such conditions by making physiological adjustments to facilitate greater heat dissipation. These include increasing evaporative water loss or undertaking facultative hyperthermia [12]. Alternatively or concurrently, they may make behavioural adjustments to lower heat load, including reducing activity and shifting into the shade [13]. Such adjustments should be non-linear in nature, as energy and water requirements for thermoregulation vary little when ambient temperatures fall within an endotherm's thermoneutral zone (TNZ), but increase dramatically outside of it [12,14,15]. Furthermore, evaporative cooling

becomes the only means of heat-dissipation when air temperature exceeds body temperature [15]. This results in sharp increases in water loss with associated danger of dehydration, especially for small animals [16,17]. Thermoregulation at temperatures above the TNZ therefore carries high water and energy costs, potentially accompanied by reduced intake [7,18] and missed-opportunity costs inherent in behavioural adjustments.

High temperatures can influence survival directly, or can have more subtle effects on fitness. For example, under high temperature conditions in the Kalahari, southern pied babblers (*Turdoides bicolor*) traded-off heat dissipation behaviours against foraging efficiency. This led to reduced mass gain, with potential implications for fitness [18]. Exposure to high temperatures also reduced long-term survival in desert-dwelling banner-tailed kangaroo rats [19]. Arid-zones globally are predicted to experience increasing frequencies and duration of hot-weather events under climate change [20,21], potentially exposing animal populations to chronic and perhaps cumulative fitness costs [7]. Indeed, rising temperatures in Mexico have already driven declines and local extinctions of *Sceloporus* lizard populations via sub-lethal effects on activity patterns [3].

Sub-lethal costs of high temperatures may be particularly acute during breeding in species with appreciable levels of parental care.

This is because adults carry the cost of providing for dependent offspring as well as for themselves, resulting in conflicts during times of resource bottlenecks (e.g. when temperature reduces the ability of parents to forage via imposing thermoregulatory costs or reducing prey availability). Reduced provisioning rates to nests during periods of high temperatures have been documented in a number of arid-zone passerine species [22–24]. High temperatures may also reduce growth rates in young birds under conditions of *ad libitum* food availability [25], possibly due to increased thermoregulatory demands. Nestlings therefore face a double challenge during hot weather: coping with reduced parental care at a time when their own thermoregulatory costs are high.

We studied effects of high daily maximum air temperature ( $t_{\max}$ ) on aspects of breeding success in a species near the ‘warm edge’ of its range: the common fiscal (*Lanius collaris*) in the southern Kalahari. The common fiscal is widespread in sub-Saharan Africa including the periphery of the Kalahari, but largely absent in the central Kalahari basin [26]. It is a typical Laniidae shrike, hunting invertebrates and small vertebrates from exposed perches [27], leaving it vulnerable to high levels of solar radiation. We hypothesised that higher  $t_{\max}$  would therefore negatively affect nest provisioning rates. Reduced provisioning may in turn result in lower rates of body mass gain by nestlings, with implications for size and age at fledging. As air temperatures are likely to correlate with nest temperatures (shown to affect nestling growth in other birds e.g. [28]), we further hypothesised that  $t_{\max}$  would affect nestling body mass directly. We examined these effects on broods at different stages, because the effect of temperature can be modified by nestling age [28,29]. Due to the non-linear nature of physiological responses to temperature, we explored the existence of  $t_{\max}$  thresholds above which the size of effects on fledging parameters began to increase. Finally, we calculated the implications of delayed fledging for nestling survival in terms of increased vulnerability to time-dependant mortality [30].

Our aim was to provide data on links between climate, specifically temperature, and breeding success for a common bird with a wide geographical range. This data can inform predictions of how absolute fitness and population persistence may change under scenarios of global warming [7]. We used  $t_{\max}$  as our standard measure of temperature because (a) we were interested in the effects of temperature at the scale of days, (b)  $t_{\max}$  is likely to correlate with the range of environmental temperatures (*sensu* [31]) available to birds and can be used as an index of such (e.g. see [3,18]), and (c)  $t_{\max}$  is collected at weather stations globally and is commonly used in climate change analyses (e.g. [32]).

## Materials and Methods

### Ethics statement

The methods used in this study were approved by the University of Cape Town Animal Ethics Committee (clearance # 2011/V21/PH). The study was carried out on private land (Tswalu Kalahari Reserve) with permission of the landowners and of the Northern Cape Conservancy of South Africa (permit # Fauna 1088/2011). Bird banding was done by individuals licensed by the South African Bird Ringing Unit (SAFRING) and all bird handling was done by experienced individuals.

### Study site

We worked in a 10 km<sup>2</sup> area of dunefield on Tswalu Kalahari Reserve (100 000 ha; 27°13'S, 22°22'E), South Africa. Vegetation was typical of Kalahari arid savanna, consisting of sparse grasses (*Eragrostis spp.*, *Panicum spp.*, *Aristida spp.*) dotted with trees and

shrubs (*Acacia erioloba*, *A.haematoxylin*, *A. mellifera*, *Boscia albitrunca*, *Terminalia sericea*, *Rhigozum trichotomum*).

The study was done over two austral summers: November 2010 – March 2011 and November 2011 – February 2012, which corresponds with the breeding season for common fiscals. Air temperature maximum and minimum were 38.7°C and 8.7°C in 2010/11 and 39.1°C and 9.3°C in 2011/12, respectively. Average daily rainfall was 4 mm in 2010/11 and 3 mm in 2011/12. Rain fall was erratic with most falling heavily during thunderstorms interspersed with long dry periods. Meteorological data were collected using an onsite weather station (VantagePro 2, Davis Instruments, Hayward, California).

### Study species and population

The common fiscal is a medium-sized passerine (35–45 g) endemic to Africa. The Kalahari population is often considered a subspecies (*L. c. subcoronatus*) but recent genetic analysis does not support this distinction [33]. Common fiscals are territorial, socially monogamous, open cup breeders [27]. Pairs defended territories of ~3 – 10 ha at our study site.

We captured fifty-four individuals (28 males, 26 females) using springtraps baited with giant mealworms (*Zophobas morio*). We sexed the birds using presence/absence of a rufous flank patch (present only in females), and fitted them with three plastic colour bands (JC Hughes, England) and one uniquely numbered aluminium or Incoloy band for individual identification. We monitored 21 breeding pairs of colour-banded individuals (in 19 pairs both partners were banded, two pairs contained an un-banded female we were unable to catch). Fifteen of these pairs produced broods which survived > 6 days post-hatch and therefore contributed to our analyses, and four of these pairs contributed two broods each (19 broods in total). Modal brood size was three nestlings (range: 1–4 nestlings). The remaining 14 colour-banded birds were not recorded breeding during the study.

We visited territories on alternate days to ascertain breeding status of the pair, and once nests were initiated, to establish lay and hatch dates of eggs and survival of nestlings. Nestlings hatched asynchronously over one – three days. We made three full-day observations at each nest when the first-hatched nestling was six (n = 17 nests), ten (n = 13 nests), and 14 (n = 12 nests) days old, taking day of hatch as day one (of the 13 nests that survived to fledge, 10 were sampled at all three ages). We banded all nestlings with an aluminium or Incoloy numbered ring either on the evening of day 14 or the morning of day 15.

### Nestling mass gain

We weighed nestlings on a top-pan balance (DS50, Pesola, Baar, Switzerland) twice on each observation day at approximately 6h00 and 18h00. Weighing sessions took ~ 1 min/nestling after which we immediately returned nestlings to the nest. We calculated diurnal change in body mass ( $\Delta_m$ ) as a percentage of morning body mass and standardised it for minor variations in timing of morning and evening weighing sessions using the following formula (from [18]):

$$\Delta_m = 100((w_2 - w_1)/w_1)/((t_2 - t_1)/12)$$

where  $t_2 - t_1$  is the number of decimal hours between morning ( $t_1$ : ~6h00) and evening ( $t_2$ : ~18h00) weighing sessions;  $w_1$  is mass in morning, and  $w_2$  is mass in evening.

We marked nestlings on the tarsus or toes with nail varnish for individual identification during the morning weighing session. These markers lasted >12 hours but disappeared between

measurement days. This meant we could collect mass change data for individuals at the scale of one day, but could not identify these individuals again on the next observation day. We therefore analysed data as averages per brood to avoid pseudo-replication.

### Provisioning rates

We placed video cameras (Sony HDR-XR160E; Panasonic SDR-S50) on a tripod 2–5 m from the nest tree immediately after the morning weighing session on observation days (~6h00) and retrieved them immediately before the evening weighing session (~18h00). We extracted data on provisioning rates to nestlings from videos. We divided total number of provisions by number of video-recording hours, then multiplied by 12 to standardise provisioning rate to a 12 hr day. We discarded observations where video recording length was < 10 hrs, which occasionally occurred due to equipment failure or rainstorms, to avoid introducing time-of-day biases into data.

### Fledging mass, tarsus length and age-at-fledge

We used body mass and tarsus length (measured with Vernier callipers) measurements taken when banding the nestlings as a proxy for fledging body mass and tarsus length [34,35]. We monitored nests daily after nestlings were banded to establish fledging date.

### Statistical analysis

All analyses were conducted in the R statistical environment [36] using packages lme4 version 0.999375-35 [37], lsmeans version 1.06-05 [38], and MuMIn version 0.13.14, [39].

Factors influencing total provisions per brood per 12-hr day,  $\Delta m$ , and age, tarsus length and body mass at fledging, were investigated by fitting Generalized Linear Mixed Models (GLMMs). Residuals of global and final models were visually inspected to ensure model assumptions were met. Fits of all possible nested models for each analysis were compared using AICc (Akaike's Information Criteria, adjusted for small samples); models were considered better if they reduced AICc by > 2.

**Total provisions per day.** We analysed provisioning data using a GLMM with Poisson error structure and a log-link function. Four pairs contributed two broods each, but models including the random term brood identity nested within pair identity failed to converge. We therefore removed all observations of one brood each (the brood for which we had fewer observations or, if the number of observations were equal, selected at random) for pairs which had contributed two broods to the dataset and refitted the model including only the random term brood identity.

We included the following fixed factors in the global model  $t_{max}$ , brood size, nestling age, and all two-way interactions. We removed a single observation with unusually high leverage from the analysis, but this did not influence parameters included in the final model.

Model predictions and 95% CIs were back-transformed by taking the exponential of the sum of the model prediction and the variance component for the random term.

**Nestling  $\Delta m$ .** We analysed nestling  $\Delta m$  data using a GLMM with Gaussian error structure and an identity-link function. Brood size and total provisions per day were strongly correlated. To avoid issues of collinearity, we fitted total provisions per day adjusted for brood size (provisions per nestling = provisions per day/brood size).

The global model included the random term brood identity nested within pair identity, and fixed factors  $t_{max}$ , provisions per nestling, nestling age and all two-way interactions. Two observa-

tions were removed due to unusually high leverage; this did not influence parameters included in the final model.

**GLMMs for fledging parameters.** We used three approaches to assess the relationship between hot weather during the nestling period and fledging size and age. Firstly, we explored the relationship between mean  $t_{max}$  during the nestling period and (a) fledging body mass, (b) fledging tarsus length, or (c) age-at-fledge by fitting separate GLMMs for each response. Each model contained the sole fixed factor mean  $t_{max}$  during the nestling period. Models for body mass and tarsus length contained the random term brood identity nested within pair identity. Age-at-fledge models contained one data point per brood (as all nestlings within each brood fledged on the same day) and were fitted with the random term pair identity.

Secondly, we explored whether critical threshold  $t_{max}$ s ( $t_{crit}$ s) existed. To do this, we investigated whether increasing numbers of days during the nestling period on which  $t_{max} > t_{crit}$  would affect fledging size and age. For each fledging parameter we fitted a separate model for each 1°C increment in  $t_{max}$  from 27°C to 38°C (candidate  $t_{crit}$ s). We used an identical model structure to that described above, but replaced the fixed effect *mean  $t_{max}$*  with the *number of days  $t_{max} > t_{crit}$* . Effect sizes and 95% CIs for each model were then plotted against candidate  $t_{crit}$ s to show trends in strength and direction of relationships. This allowed identification of threshold  $t_{crit}$ s above which fledging parameters were compromised.

Finally, we explored whether fledglings were more vulnerable to hot weather at specific stages of the nestling period, by modelling the same fledging parameters as a function of random terms described above and  $t_{max}$  on each day of the nestling period. Effect sizes and 95% CIs were plotted against each day during the nestling period to highlight stages at which high  $t_{max}$  most influenced fledging mass, tarsus length, and age. All GLMMs for fledging parameters had a Gaussian error structure and identity link function.

**Survival analysis.** We used the Mayfield estimator [40] to estimate daily survival probability of nests once eggs had hatched:

Daily survival probability = 1 - (number of failed nests/total number of days survived by all nests)

We estimated daily failure risk as:

Daily failure risk = 1 - daily survival probability

All data are presented as means (95% CIs), unless otherwise stated. We opted not to report statistical significance (p-values) in order to focus attention on biological relevance of effect sizes, following Garamszegi et al. [41], Nakagawa and Cuthill [42], and others.

## Results

### Daily provisioning rate

We had only one best-fit daily provisioning rate model, which had a model weight of 0.819 (Table 1). The sample size for this analysis was 28 observations of 12 broods from 12 pairs.

The best-fit model contained nestling age, brood size,  $t_{max}$ , and brood size\* $t_{max}$ , as well as the random factor brood identity (Table 2). Daily provisioning rates to broods of ten-day old nestlings (mean: 84.9; range: 72.2–100.1) were higher on average than to broods of six- (68.2; 57.4–81.0) or 14-day old (65.7; 55.0–78.4) nestlings.

There was a negative brood size\* $t_{max}$  interaction, such that larger broods experienced a greater reduction in total provisions per day than smaller broods as  $t_{max}$  increased. Visual examination of data suggested this was because larger broods received greater

**Table 1.** Top five models for total daily provisioning rate.

| Model   | k | Dev   | AICc  | ΔAICc | Model weight |
|---|---|-------|-------|-------|--------------|
| Nestling age + brood size + $t_{max}$ + brood size* $t_{max}$                           | 7 | 68.97 | 88.57 | 0.000 | 0.819        |
| Nestling age + $t_{max}$  | 5 | 81.14 | 93.86 | 5.293 | 0.058        |
| Nestling age + brood size + $t_{max}$ + nestling age* $t_{max}$ + brood size* $t_{max}$ | 9 | 66.59 | 94.59 | 6.019 | 0.040        |
| Nestling age + brood size + $t_{max}$   | 6 | 78.99 | 94.99 | 6.419 | 0.033        |
| Nestling age + $t_{max}$ + nestling age* $t_{max}$                                      | 7 | 76.79 | 96.39 | 7.819 | 0.016        |

\*Dev = model deviance. Global model: nestling age + brood size +  $t_{max}$  + nestling age\*  $t_{max}$  + brood size\*  $t_{max}$  + nestling age\*brood size. Random term: brood identity. N = 28 observations of 12 broods from 12 pairs.  
doi:10.1371/journal.pone.0074613.t001

numbers of provisions than smaller broods at low  $t_{max}$ , but at high  $t_{max}$  brood size no longer influenced provisioning rate.

### Nestling mass change

We had only one best-fit model for nestling  $\Delta m$ , which had a model weight of 0.953 (Table 3). The sample size for this analysis was 38 observations of 18 broods from 14 pairs. This model contained nestling age, provisions per nestling,  $t_{max}$ , and the interaction between  $t_{max}$  and nestling age, as well as the random effect of brood identity nested within pair identity (Table 4). There was a strong negative effect of increasing  $t_{max}$  on nestling  $\Delta m$  for six-day old nestlings (where 1°C increase in  $t_{max}$  resulted in 2.5% less body mass gain), but increasing  $t_{max}$  had a negligible effect on older nestlings (1°C increase in  $t_{max}$  resulted in < 0.1% reduction in body mass gain, Fig. 1, Table 4). Increasing numbers of provisions per nestling positively influenced  $\Delta m$ , with each additional provision increasing body mass gain by 0.36% (Table 4). Daily body mass gain of six-day old nestlings (27.5%; 22.1 – 32.9%) was higher on average than ten (13.2%; 9.5 – 16.8%) or 14-day old nestlings (4.3%; 2.0 – 6.5%). In absolute terms, these % body mass gains were equivalent to 2.7 g (2.2 – 3.2 g) for six-day old chicks, 2.7 g (2.0 – 3.3 g) for ten-day old chicks, and 1.2 g (0.5 – 1.9 g) for 14-day old chicks (average morning body mass was 10.8 g (9.6 – 12 g), 22.7 g (20.7 – 24.7 g) and 29.3 g (27.3 – 31.2 g), respectively).

**Table 2.** Factors affecting total daily provisioning rate, estimates of effect sizes, standard errors (SE), and 95% confidence intervals (95% CI).

| Variable               | Estimate | SE   | 95% CI        |
|------------------------|----------|------|---------------|
| $t_{max}$              | 0.05     | 0.03 | -0.01 – 0.11  |
| Brood size             | 1.30     | 0.36 | 0.59 – 2.01   |
| Nestling age:          |          |      |               |
| six days               | 2.14     | 1.01 | 0.16 – 4.12   |
| ten days               | 2.36     | 0.99 | 0.42 – 4.30   |
| fourteen days          | 2.10     | 0.99 | 0.16 – 4.04   |
| Brood size * $t_{max}$ | -0.03    | 0.01 | -0.05 – -0.01 |

\*N = 28 observations of 12 broods from 12 pairs. The model was run with a Poisson error structure and log-link function. Effect size estimates are not back-transformed, therefore no units are presented.  
doi:10.1371/journal.pone.0074613.t002

### Relationships between hot weather during the nestling period and fledging parameters

**Mean  $t_{max}$ .** Mean  $t_{max}$  during the nestling period had negative effects on fledging body mass (-1.33 g per 1°C increase in mean  $t_{max}$ , 95% CI: -2.9 – 0.25) and tarsus length (-0.10 mm per 1°C increase in mean  $t_{max}$ , 95% CI: -0.69 – 0.49), but 95% CIs were large and included zero in both cases. Modal age-at-fledge was 18 days (range: day 15 to day 21). Mean  $t_{max}$  had a positive effect on age-at-fledge (+ 0.70 days per 1°C increase in mean  $t_{max}$ ), but again the 95% CI (-0.31 – 1.71) was large and included zero. Direction of the effect of mean  $t_{max}$  on fledging parameters was therefore uncertain, and wide CIs suggest this measure does not capture well the mechanisms underlying variation in fledging parameters.

**Effects of exceeding  $t_{max}$  thresholds.** Increasing frequency with which  $t_{max} > t_{crit}$  (see Methods) during the nestling period resulted in reduced fledging weight and tarsus length as well as delaying fledging. Threshold  $t_{crit}$ s exist for all parameters, above which the strength of relationships increased (Fig. 2). For example, the number of days during the nestling period on which  $t_{max}$  exceeded  $t_{crit} \geq 33^\circ\text{C}$  had negative effects on fledging body mass (90 or 95% CIs for these estimates do not contain zero) which increased in strength as  $t_{crit}$  increased (Fig. 2A).

**Impact of  $t_{max}$  at different nestling stages.** Days with high  $t_{max}$  early in the nestling period tended to promote higher body mass and longer tarsi in fledglings, and earlier fledging (Fig. 3). These effects were reversed during the middle of the nestling period, when high  $t_{max}$  tended to delay fledging and reduce fledging body mass and tarsus length. High  $t_{max}$  had a particularly strong negative effect on fledging body mass on days eight, nine, and ten (Fig. 3A), and on fledging tarsus length on days eight and nine (Fig. 3B). Late in the nestling period (days eleven and following) all  $t_{max}$  effects diminished in size (Fig. 3). For all fledging parameter models, there were 32 fledglings from 13 nests.

### The risks of delaying fledging

Forty-five nests were initiated by colour-banded birds during this study. Of these, twenty-five survived to hatching, and thirteen survived to fledge. The daily post-hatching survival probability of nests was 95.8%. Therefore, risk of nest failure increased by 4.2% per additional day spent in the nest, post-hatching.

### Discussion

Increasing numbers of hot days during the nestling period affected quality, and potentially quantity, of common fiscal fledglings produced, suggesting a mechanism by which predicted temperature increases in the Kalahari [21,43] could negatively

**Table 3.** Top five models for nestling % daily body mass gain ( $\Delta m$ ).

| Model   | k  | Dev   | AICc  | $\Delta AICc$ | Model weight |
|---|----|-------|-------|---------------|--------------|
| Nestling age + ppn + $t_{max}$ + nestling age* $t_{max}$                    | 10 | 216.0 | 244.1 | 0.000         | 0.953        |
| Nestling age + ppn + $t_{max}$ + nestling age*ppn + nestling age* $t_{max}$ | 12 | 214.9 | 251.4 | 7.305         | 0.025        |
| Nestling age + ppn + $t_{max}$ + nestling age* $t_{max}$ + ppn* $t_{max}$   | 11 | 220.8 | 253.0 | 8.840         | 0.011        |
| Nestling age + ppn + $t_{max}$  | 8  | 233.4 | 254.4 | 10.270        | 0.006        |
| Nestling age + $t_{max}$ + nestling age* $t_{max}$                          | 9  | 230.9 | 255.4 | 11.260        | 0.003        |

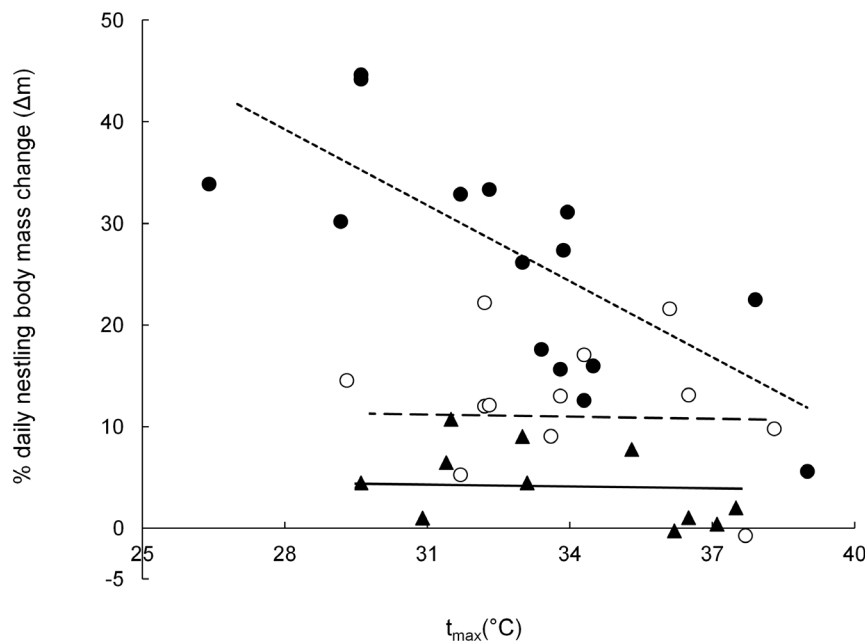
\*ppn = provisions per nestling. Dev = model deviance. Global model: nestling age + ppn +  $t_{max}$  + nestling age\*  $t_{max}$  + ppn\*  $t_{max}$  + nestling age\*ppn. Random term: brood identity nested within pair identity. N = 38 observations of 18 broods from 14 pairs.  
doi:10.1371/journal.pone.0074613.t003

affect populations. We found a strong negative effect of increasing  $t_{max}$  on the daily body mass gain of young (six-day old) nestlings. This effect was mediated by reduced parental provisioning rates to larger broods on hotter days, and additionally by direct effects of high temperatures on nestlings. We found only weak evidence for a relationship between mean  $t_{max}$  over the entire nestling period, and fledgling body mass, tarsus length and age-at-fledge. Instead, these factors were strongly influenced by frequency with which  $t_{max}$  exceeded certain thresholds ( $t_{crit}$ s). Increasing numbers of days of  $t_{max} > 33^\circ\text{C}$  during the nestling period resulted in reduced fledgling body mass, increasing number of days of  $t_{max} > 37^\circ\text{C}$  negatively influenced fledgling tarsus length, and increasing number of days of  $t_{max} > 35^\circ\text{C}$  positively influenced age-at-fledge. Furthermore, the impact of hot weather on fledgling parameters was greatest when high  $t_{max}$  days occurred during the middle of the nestling period ( $\sim$ days 7–10). Thus, ‘hot’ nestling periods, measured in terms of number and timing of hot days (as opposed to the average  $t_{max}$ ), resulted in smaller fledglings which

left the nest later. This had implications both for probability of fledging (daily time-dependant mortality risk of nestlings was 4.2%) and potentially for fledgling lifetime fitness [44–46].

### Mechanisms underpinning temperature effects on daily nestling body mass gain

Effects of environmental variables (weather conditions, habitat quality) on nestling growth are often mediated by parental provisioning rates [47–49]. We found a negative relationship between  $t_{max}$  and provisioning rate in common ferals that interacted with brood size, such that large broods experienced the greatest reduction in provisioning as temperatures increased. Much of the prey provisioned by common ferals in our study was arthropod invertebrates. Like other ectotherms, invertebrates make behavioural adjustments to regulate body temperature within an optimum range [50]. In cool climates, avian provisioning rates may therefore increase with increasing temperature, as invertebrates become more active [51,52]. However, above an



**Figure 1. The effect of maximum daily temperature on % daily mass change of nestlings.** The negative effect of  $t_{max}$  on % daily mass change of nestlings ( $\Delta m$ ) was modified by nestling age. Closed circles and dotted line represent six-day old nestlings; open circles and dashed line represent ten-day old nestlings, stars and solid line represent 14-day old nestlings. Each data point represents  $\Delta m$  averaged over all nestlings in a single brood. Lines of best fit were calculated using model predictions from a GLMM at average levels of provisioning and take into account variance caused by random terms.  
doi:10.1371/journal.pone.0074613.g001

**Table 4.** Factors affecting nestling % daily mass change ( $\Delta m$ ), estimates of effect sizes, standard errors (SE), and 95% confidence intervals (95% CI).

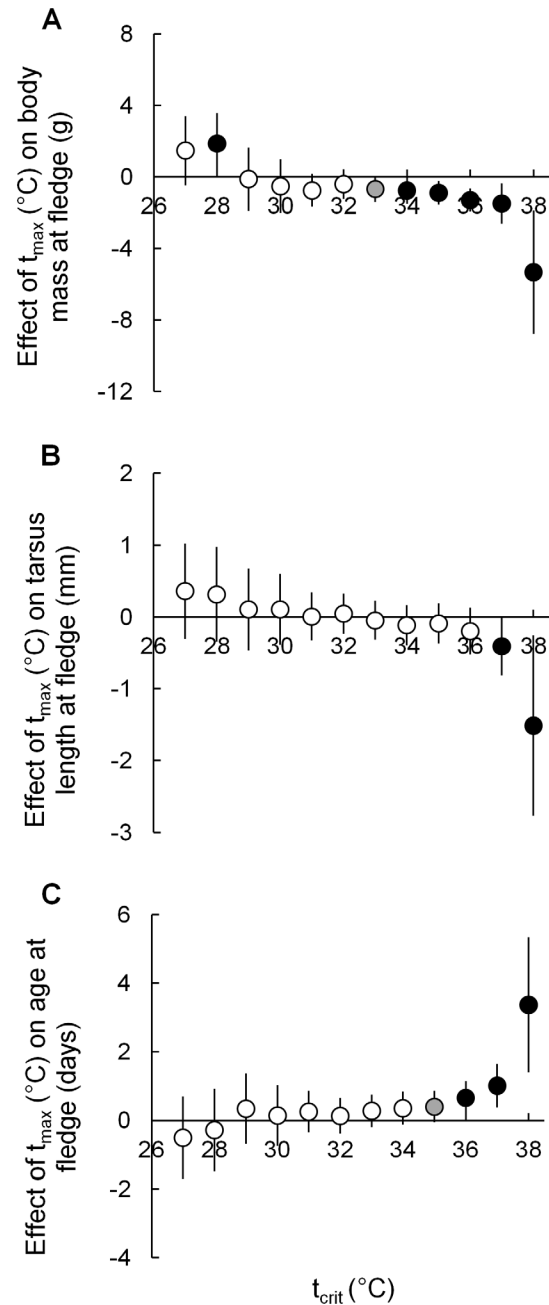
| Variable                 | Estimate | SE    | 95% CI         |
|--------------------------|----------|-------|----------------|
| $t_{max}$                | -0.06    | 0.56  | -1.16 – 1.04   |
| Nestling age:            |          |       |                |
| six days                 | 97.69    | 13.42 | 71.39 – 123.99 |
| ten days                 | 2.10     | 19.57 | -36.26 – 40.46 |
| fourteen days            | -5.13    | 19.78 | -43.90 – 33.64 |
| Provisions per nestling  | 0.36     | 0.07  | 0.22 – 0.50    |
| $t_{max}$ *nestling age: |          |       |                |
| $t_{max}$ *six days      | -2.49    | 0.40  | -3.27 – -1.71  |
| $t_{max}$ *ten days      | -0.07    | 0.56  | -1.17 – 1.03   |
| $t_{max}$ *fourteen days | -0.06    | 0.56  | -1.16 – 1.04   |

\*N = 38 observations of 18 broods from 14 pairs. Units for estimates of effect size are % daily body mass change ( $\Delta m$ ).  
doi:10.1371/journal.pone.0074613.t004

upper temperature threshold, invertebrates seek out cooler microclimates, retreating underground or into shade. Therefore, reductions in common fiscal provisioning rates on hot days may have been due to reduced prey availability. Alternatively, fiscals may have traded-off provisioning behaviour against their own thermoregulatory requirements, as documented in other desert-dwelling passerines [22,23]. Further work is needed to assess which of these processes is most important in determining provisioning rates of insectivorous birds during periods of high temperature.

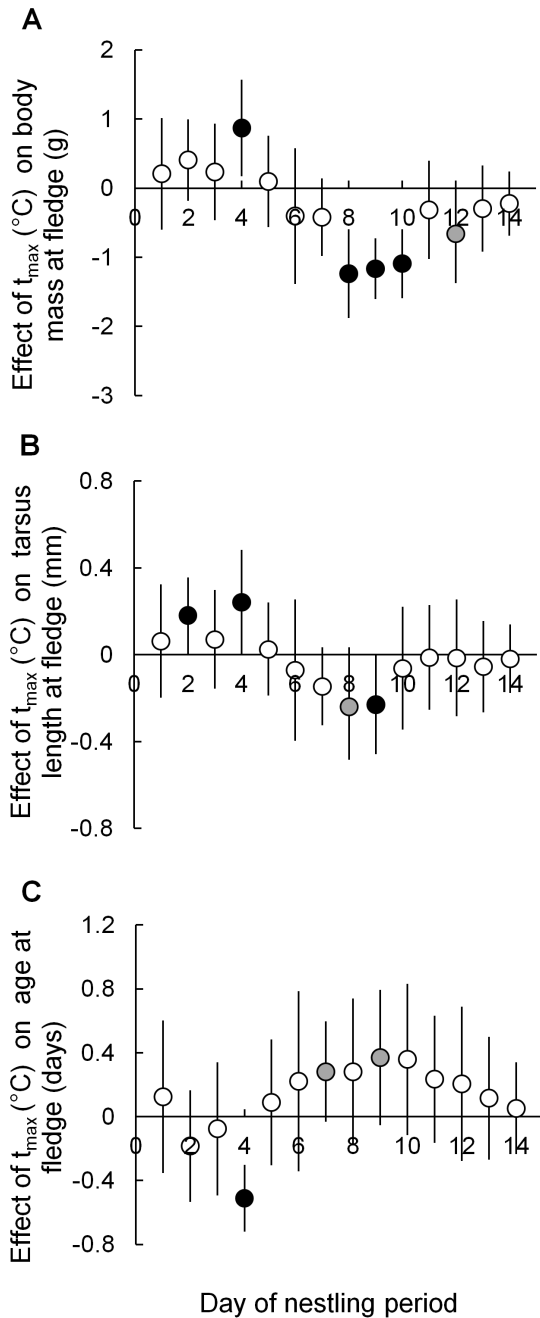
In our study, daily nestling mass gain declined on hot days to a greater extent than could be explained by reductions in provisioning rate alone, suggesting that direct physiological costs of high  $t_{max}$  (extra expenditure of energy and water or reduction in the efficiency of physiological processes [15]), may also play a role. We were unable to quantify precisely the relative importance of reduced provisioning vs. direct temperature effects on daily nestling mass gain, for two main reasons. Firstly, we were able to quantify the number of provisions brought to the nest but not biomass delivered. This was due to low quality video from standard definition cameras and lack of data on biomasses of identifiable prey items. Breeding common fiscals are central place foragers (with the nest as the ‘central place’). Central Place Foraging Theory predicts that when costs of foraging are high, parents should return with larger loads [53]. Despite this, conflicting evidence exists to suggest that some avian parents may actually reduce biomass of food loads brought to the nest when foraging is costly [47,54,55]. It is therefore uncertain whether our assessment of only provisioning rate and not biomass was more likely to have underestimated or overestimated reductions in food delivery at high temperatures.

Secondly, once provisioning rate was accounted for, the effect of temperature on nestling mass gain was modified by nestling age. Six-day old common fiscals showed stronger reductions in body mass gain on hot days than ten- or 14-day old nestlings. Similar patterns have been shown in other species [28,29], suggesting that younger nestlings may generally be more vulnerable to temperature effects. Furthermore, six-day old nestlings in our study gained a much higher percentage of their body mass per day on average, than ten- or 14-day old birds, perhaps making temperature-related effects on mass gain easier to detect at this age (measurement errors would be smaller relative to the effect



**Figure 2.** Relationships between numbers of days  $t_{max} > t_{crit}$  during the nestling period and fledging parameters. Model estimates and 95% CIs for relationships between number of days on which  $t_{max} > t_{crit}$  (plotted on the x-axis) and A: fledging mass; B: fledging tarsus length; and C: age-at-fledge. The y-axis shows the effect on each fledging parameter of a single day during the nestling period on which  $t_{max} > t_{crit}$ . For example in A: a single day of  $t_{max} > 38^\circ\text{C}$  will reduce fledgling body mass by 5.3 g. As number and intensity of hot days within the nestling period increases, the size of fledglings decreases (A & B) and nestlings take longer to fledge (C). White circles indicate 90% CIs include zero; grey circles indicate that 90% CIs exclude zero; black circles indicate that 95% CIs exclude zero.  
doi:10.1371/journal.pone.0074613.g002

size). Interestingly, there was no interaction between age and provisioning rate on nestling mass gain. This suggests the relative importance of *direct* temperature costs (as opposed to those realised through provisioning) were higher for younger than older



**Figure 3. Relationships between  $t_{max}$  on each day of the nestling period and fledging parameters.** Model estimates and 95% CIs for relationships between  $t_{max}$  on each day of the nestling period (day 1 = day of hatch) and A: fledging mass (g); B: fledging tarsus length (mm); and C: age-at-fledge. The y-axis shows the effect of a 1°C increase in  $t_{max}$  on the day of the nestling period indicated on the x-axis. For example in A: a 1°C increase in  $t_{max}$  on day eight of the nestling period reduced fledging body mass by 1.2 g. White circles indicate 90% CIs include zero; grey circles indicate that 90% CIs exclude zero; black circles indicate that 95% CIs exclude zero. doi:10.1371/journal.pone.0074613.g003

nestlings. Studies in controlled environments where provisioning rates and ambient temperature can be varied independently (see [25,56]) might help to better disentangle these effects.

### Impact of temperature on size and age at fledging

Contrary to expectations, strong effects of high  $t_{max}$  on body mass gain of six-day old nestlings were not reflected in their fledging body mass. Instead, high temperatures later in the nestling period (particularly days eight and nine; Fig. 3), had a greater effect on body mass, tarsus length, and age-at-fledge. Growth of avian nestlings usually approximates a sigmoidal curve, with the most rapid growth rates occurring mid-nestling period [46]. Our data suggest the period of most rapid growth had commenced by day six, growth rates were declining by day ten, and nestlings were approaching asymptotic body mass by day 14. Organisms may compensate for suppressed growth during early development by accelerating growth rates when conditions improve, although this can carry costs later in life (reviewed by [57]). Younger nestlings may therefore have been able to compensate for negative effects of high temperatures, both in terms of growth and recovery from dehydration, whereas older nestlings may have had insufficient time to “catch up” prior to fledging.

Lack of strong evidence for an effect of mean  $t_{max}$  on fledging size is unsurprising given the differential impacts of high  $t_{max}$  on different days during the nestling period. Instead, fledging parameters were affected by the frequency during the nestling period with which  $t_{max}$  exceeded critical thresholds. This importance of thresholds, as opposed to mean temperatures, is likely due to the non-linear shape of physiological responses to temperature [15]. For example, for adult common fiscals the upper critical limit of the TNZ is between 35°C and 38°C ambient temperature. Above this critical limit, metabolic expenditure related to thermoregulation increases dramatically. Common fiscals use facultative hyperthermia (raise the body temperature ~2°C above normal) to reduce water costs of thermoregulation at ambient temperatures > 30°C in the lab [58]. To the best of our knowledge, the physiological response of common fiscal nestlings to elevated temperature has never been studied; therefore the upper critical limit of the TNZ for a homeothermic fiscal nestling is unknown. Despite this, we can assume such a threshold exists and may be implicated in the non-linear relationship between nestling-period  $t_{max}$ s and fledging outcomes we observed.

The relationship between ambient temperature in the lab, air temperature in the field and range of environmental temperatures experienced by birds is complex due to the influence of wind, solar radiation, and humidity, and variation in physical characteristics of the birds themselves [59]. However, in absence of wind, environmental temperature in the shade may approximate air temperature [60]. Daily  $t_{max}$  probably therefore represents one of the cooler thermal environments available to adult common fiscals at the hottest time of day (on windy days, increased convective heat loss may mean standard operative environmental temperatures (*sensu* [31]) in the shade could be even cooler than air temperature). In this study, environmental temperatures in nests were likely higher than air temperature as many were at least partially exposed to the sun. Despite these complexities, the range of environmental temperatures available to fiscals at the hottest time of day is likely to correlate, to some degree, with  $t_{max}$ . (i.e. environmental temperatures are likely to be higher overall on high  $t_{max}$  days). On hot days with  $t_{max}$  near the upper critical limits measured in the lab, fiscals would be likely to encounter a range of environmental temperatures including some above their TNZ. Thermoregulatory costs at such environmental temperatures perhaps played a role in observed reductions in provisioning rates by adult birds. Under such conditions, homeothermic nestlings may also have been obliged to channel resources towards thermoregulation that might otherwise have been used for growth.

## Implications for body size patterns

Thermoregulatory considerations underlie the predictions of Bergmann's Rule that endotherms should become larger with increasing latitude because a lower surface area to volume ratio helps conserve body heat [61]. In hot environments, smaller body size may therefore confer thermoregulatory advantages through increasing efficiency of convective cooling - provided environmental temperatures remain below body temperature. Interestingly, a recently study of museum specimens documented reductions in body size in Australian passerines in accordance with increasing temperatures over the last century [62], reflecting similar trends observed in Israel [63]. Gardner *et al.* [62] found no ptilochronological evidence of nutritional stress as a driver of the body size reductions, and therefore suggested they might be an adaptive response to climate change, in keeping with Bergmann's Rule [62,64]. The effect of temperature we found on fledging size is unlikely to be detectable in feather growth bars of adults after their first post-fledging moult. We therefore suggest a third potential explanation: direct (physiological) and indirect (nutritional) effects of higher temperatures during critical stages of nestling growth, which translate into reductions in adult body size.

## Implications for individual fitness and populations under warming trends

Reduced fledging size in birds is often correlated with reduced survival, recruitment into the breeding population, fecundity, and reproductive success [45,65–68]; however see [69,70]. Temperature-driven variation in body size is therefore likely to have far-reaching implications for life-histories of common ferals that experience 'hot' weather conditions while in the nest. The southern Kalahari and surrounding areas in north-western South Africa are undergoing among the fastest rates of warming in the region [32], with the implication that such 'hot' nestling periods will become more frequent. This is especially the case because arid-zone birds tend to breed in response to the onset of rainfall [71] and are therefore unlikely to be able to advance breeding dates to avoid the hottest part of the season. As warming continues, high temperatures during the breeding season may

affect the absolute fitness of increasing proportions of individuals within each cohort of fledglings, with consequences for the maintenance of the southern Kalahari feral population.

We believe that investigations of the shape of the relationship between  $t_{max}$  and correlates of fitness, such as presented here, are of great importance for predicting species' responses to climate change. In this study, we discovered biologically meaningful critical threshold  $t_{max}$ s which if exceeded during the nestling period, are likely to affect the fitness of common feral fledglings. It is probable that such thresholds exist in other species as well, due to the non-linear nature of physiological responses to temperature. Identification of these thresholds will provide solid data for use in predicting the impacts of past and future climate change on populations and species.

## Acknowledgments

Laura Barclay, Rob Sutcliffe, David Nkosi, Ben Smit, Sue-Joy and Barry Schultz, and Christine Moore provided field assistance, Rob Simmons loaned us field equipment, and Anthony Schultz, Yolanda Chirango, Jess Berndt, and Bruce Baigrie helped with data transcription. Tim Reid, Thomas Flower, and Arjun Amar provided insight into statistical modelling techniques in R, and Alex Thompson and Warwick Hojem commented on the manuscript. Duncan MacFadyen and E. Oppenheimer & Sons via the Tswalu Foundation gave permission to work on Tswalu Kalahari Reserve and Gus van Dyk and Dylan Smith facilitated work there. We are also grateful for further logistical support provided by Tswalu staff members. We thank two anonymous reviewers for comments and insight which improved the manuscript. We also thank Justin Boyles for his careful editing which improved the clarity of the manuscript. SJC, ROM, and CLH would also like to use this opportunity to remember PARH, who sadly passed away while this manuscript was still in preparation.

## Author Contributions

Conceived and designed the experiments: SJC ROM PARH. Performed the experiments: SJC CLH. Analyzed the data: SJC ROM. Contributed reagents/materials/analysis tools: SJC ROM PARH. Wrote the paper: SJC ROM. Edited manuscript versions: SJC ROM CLH PARH. Obtained funding: PARH. Developed relationship with landowners: PARH.

## References

1. Parmesan C, Yohe G (2003) A globally coherent fingerprint of climate change impacts across natural systems. *Nature* 421: 37–42. Available: <http://www.ncbi.nlm.nih.gov/pubmed/12511946>.
2. Chen IC, Hill JK, Ohlemüller R, Roy DB, Thomas CD (2011) Rapid range shifts of species associated with high levels of climate warming. *Science* (New York, NY) 333: 1024–1026. Available: <http://www.ncbi.nlm.nih.gov/pubmed/21852500>. Accessed 28 January 2013.
3. Sinervo B, Méndez-de-la-Cruz F, Miles DB, Heulin B, Bastiaans E, et al. (2010) Erosion of lizard diversity by climate change and altered thermal niches. *Science* 328: 894–899. Available: <http://www.ncbi.nlm.nih.gov/pubmed/20466932>. Accessed 2010 July 17.
4. Cahill AE, Aiello-Lammens ME, Fisher-Reid MC, Hua X, Karanewsky CJ, et al. (2013) How does climate change cause extinction? *Proceedings of the Royal Society B: Biological Sciences* 280: 20121890. Available: <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3574421&tool=pmcentrez&rendertype=abstract>.
5. Kearney M, Porter W (2009) Mechanistic niche modelling: combining physiological and spatial data to predict species' ranges. *Ecology Letters* 12: 334–350. Available: <http://www.ncbi.nlm.nih.gov/pubmed/19292794>.
6. Buckley LB, Urban MC, Angilletta MJ, Crozier LG, Rissler LJ, et al. (2010) Can mechanism inform species' distribution models? *Ecology Letters* 13: 1041–1054. Available: <http://www.ncbi.nlm.nih.gov/pubmed/20482574>. Accessed 2013 Feb 28.
7. McKechnie AE, Hockey PAR, Wolf BO (2012) Feeling the heat: Australian landbirds and climate change. *Emu* 112: i–vii.
8. Angilletta MJ, Sears MW (2011) Coordinating theoretical and empirical efforts to understand the linkages between organisms and environments. *Integrative and Comparative Biology* 51: 653–661. Available: <http://www.ncbi.nlm.nih.gov/pubmed/21810893>. Accessed 2013 March 4.
9. Finlayson HH (1932) Heat in the interior of South Australia: holocaust of bird life. *South Australian Ornithologist* 11: 158–160.
10. Welbergen JA, Klose SM, Markus N, Eby P (2008) Climate change and the effects of temperature extremes on Australian flying-foxes. *Proceedings of the Royal Society B: Biological Sciences* 275: 419–425. Available: <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=2596826&tool=pmcentrez&rendertype=abstract>. Accessed 2013 March 25.
11. Towie N (2009) Thousands of birds die in sweltering heat. *PerthNow*. Available: <http://www.perthnow.com.au/news/thousands-of-birds-die-in-sweltering-heat/story-e6frg12c-1111118551504>. Accessed 2012 Nov 27.
12. Calder WA, King JR (1974) *Thermal and Caloric Relations of Birds*. In: Farner DS, King JR, editors. *Avian Biology*, Volume IV. New York: Academic Press. pp. 259–413.
13. Wolf BO (2000) Global warming and avian occupancy of hot deserts; a physiological and behavioral perspective. *Revista Chilena de Historia Natural* 73: 395–400.
14. Scholander PF, Hock R, Walters V, Johnson F, Irving L (1950) Heat regulation in some arctic and tropical mammals and birds. *The Biological Bulletin* 99: 237–258.
15. Angilletta MJ, Cooper BS, Schuler MS, Boyles JG (2010) The evolution of thermal physiology in endotherms. *Frontiers in Bioscience* E2: 861–881. Available: <http://www.ncbi.nlm.nih.gov/pubmed/23512519>.
16. Wolf B, Walsberg G (1996) Respiratory and cutaneous evaporative water loss at high environmental temperatures in a small bird. *The Journal of Experimental Biology* 199: 451–457. Available: <http://www.ncbi.nlm.nih.gov/pubmed/9318103>.
17. McKechnie AE, Wolf BO (2010) Climate change increases the likelihood of catastrophic avian mortality events during extreme heat waves. *Biology Letters* 6: 253–256. Available: <http://www.pubmedcentral.nih.gov/articlerender>.



- fcgi?artid=2865035&tool=pmcentrez&rendertype=abstract. Accessed 2011 Jan 17.
18. Du Plessis KL, Martin RO, Hockey PAR, Cunningham SJ, Ridley AR (2012) The costs of keeping cool in a warming world: implications of high temperatures for foraging, thermoregulation and body condition of an arid-zone bird. *Global Change Biology* 18: 3063–3070. Available: <http://doi.wiley.com/10.1111/j.1365-2486.2012.02778.x>. Accessed 2012 Nov 20.
  19. Moses MR, Frey JK, Roemer GW (2012) Elevated surface temperature depresses survival of banner-tailed kangaroo rats: will climate change cook a desert icon? *Oecologia* 168: 257–268. Available: <http://www.ncbi.nlm.nih.gov/pubmed/21833644>. Accessed 2013 March 21.
  20. Meehl GA, Tebaldi C (2004) More intense, more frequent, and longer lasting heat waves in the 21st century. *Science* 305: 994–997. Available: <http://www.ncbi.nlm.nih.gov/pubmed/15310900>. Accessed 2012 Oct 26.
  21. Akoon I, Archer E, Colvin C, Davis C, Diedericks GPJ, et al. (2011) South African Risk and Vulnerability Atlas. Pretoria: Department of Science and Technology.
  22. Austin GT (1976) Behavioral adaptations of the verdin to the desert. *The Auk* 93: 245–262.
  23. Ricklefs RE, Hainsworth FR (1968) Temperature dependent behavior of the cactus wren. *Ecology* 49: 227–233.
  24. Luck GW (2001) Variability in provisioning rates to nestlings in the cooperatively breeding Rufous Treecreeper, *Climacteris rufa*. *Emu* 101: 221–224.
  25. May JD, Lott BD (2001) Relating weight gain and feed:gain of male and female broilers to rearing temperature. *Poultry Science* 80: 581–584. Available: <http://www.ncbi.nlm.nih.gov/pubmed/11372706>.
  26. Parker V (1997) Fiscal shrike (*Lanius collaris*). In: Harrison JA, Allen DG, Underhill LG, Herremans M, Tree AJ, et al., editors. *The Atlas of Southern African Birds. Volume 2: Passerines*. Johannesburg: BirdLife South Africa. pp. 408–409.
  27. Dean WRJ (2005) Common Fiscal – *Lanius collaris*. In: Hockey PAR, Dean WRJ, Ryan PG, editors. *Birds of Southern Africa*. Cape Town: The Trustees of the John Voelcker Bird Book Fund. pp. 728–729.
  28. Murphy MT (1985) Nestling eastern kingbird growth: effects of initial size and ambient temperature. *Ecology* 66: 162–170.
  29. McCarty JP, Winkler DW (1999) Relative importance of environmental variables in determining the growth of nestling tree swallows *Tachycineta bicolor*. *Ibis* 141: 286–296.
  30. Bosque C, Bosque MT (1995) Nest predation as a selective factor in the evolution of developmental rates in altricial birds. *The American Naturalist* 145: 234–260.
  31. Bakken GS, Santee WR, Erskine DJ (1985) Operative and standard operative temperature: tools for thermal energetics studies. *American Zoologist* 25: 933–943.
  32. Kruger AC, Sekele SS (2012) Trends in extreme temperature indices in South Africa: 1962–2009. *International Journal of Climatology* 33: 661–676. Available: <http://doi.wiley.com/10.1002/joc.3455>. Accessed 2012 Nov 8.
  33. Fuchs J, Crowe TM, Bowie RCK (2011) Phylogeography of the fiscal shrike (*Lanius collaris*): a novel pattern of genetic structure across the arid zones and savannas of Africa. *Journal of Biogeography* 38: 1–13. doi:10.1111/j.1365-2699.2011.02545.x.
  34. Kilpatrick AM (2002) Variation in growth of Brown-headed Cowbird (*Molothrus ater*) nestlings and energetic impacts on their host parents. *Canadian Journal of Zoology* 80: 145–153. doi:10.1139/Z01-217.
  35. Rensel MA, Schoech SJ (2011) Repeatability of baseline and stress-induced corticosterone levels across early life stages in the Florida scrub-jay (*Aphelocoma coerulescens*). *Hormones and Behavior* 59: 497–502. Available: <http://www.ncbi.nlm.nih.gov/pubmed/21295036>. Accessed 2013 March 20.
  36. R Core Development Team. (2009) R: A language and environment for statistical computing. Available: <http://www.r-project.org>. Accessed 2011 Aug 28.
  37. Bates D, Maechler M (2010) lme4: Linear mixed-effects models using Eigen and Eigen. R package version 0.999375-35. Available: <http://cran.r-project.org/package=lme4>. Accessed 2011 Nov 17.
  38. Lenth R V (2013) lsmeans: Least-squares means. R package version 1.06-05. Available: <http://cran.r-project.org/package=lsmeans>. Accessed 2013 March 13.
  39. Barton K (2010) MuMIn: Multi-model inference. R package version 0.13.14. Available: <http://cran.r-project.org/package=MuMIn>. Accessed 2011 Nov 17.
  40. Mayfield HF (1975) Suggestions for calculating nest success. *The Wilson Bulletin* 87: 456–466.
  41. Garamszegi LZ, Calhim S, Doehrmann N, Hegyi G, Hurd PL, et al. (2009) Changing philosophies and tools for statistical inferences in behavioral ecology. *Behavioral Ecology* 20: 1363–1375. Available: <http://www.behco.oxfordjournals.org/cgi/doi/10.1093/beheco/arp137>. Accessed 2013 Jan 29.
  42. Nakagawa S, Cuthill IC (2007) Effect size, confidence interval and statistical significance: a practical guide for biologists. *Biological reviews of the Cambridge Philosophical Society* 82: 591–605. Available: <http://www.ncbi.nlm.nih.gov/pubmed/17944619>. Accessed 2012 Oct 26.
  43. Moise AF, Hudson DA (2008) Probabilistic predictions of climate change for Australia and southern Africa using the reliability ensemble average of IPCC CMIP3 model simulations. *Journal of Geophysical Research* 113: D15113. Available: <http://www.agu.org/pubs/crossref/2008/2007JD009250.shtml>. Accessed 2012 Nov 20.
  44. Lindström J (1999) Early development and fitness in birds and mammals. *Trends in Ecology & Evolution* 14: 343–348.
  45. Schwagmeyer PL, Mock DW (2008) Parental provisioning and offspring fitness: size matters. *Animal Behaviour* 75: 291–298. Available: <http://linkinghub.elsevier.com/retrieve/pii/S0003347207004253>. Accessed 2013 March 19.
  46. Gebhardt-Henrich S, Richner H (1998) Causes of Growth Variation and its Consequences for Fitness. In: Starck JM, Ricklefs RE, editors. *Avian Growth and Development*. New York: Oxford University Press. pp. 324–339.
  47. Weimerskirch H, Prince PA, Zimmermann L (2000) Chick provisioning by the yellow-nosed albatross *Diomedea chlororhynchus*: response of foraging effort to experimentally increased costs and demands. *Ibis* 142: 103–110.
  48. Schreiber EA (1994) El Niño-Southern Oscillation effects on provisioning and growth in red-tailed tropicbirds. *Colonial Waterbirds* 17: 105–119.
  49. Tremblay I, Thomas D, Blondel J, Perret P, Lambrechts MM (2004) The effect of habitat quality on foraging patterns, provisioning rate and nestling growth in Corsican Blue Tits *Parus caeruleus*. *Ibis* 147: 17–24. Available: <http://doi.wiley.com/10.1111/j.1474-919x.2004.00312.x>.
  50. Kearney M, Shine R, Porter WP (2009) The potential for behavioral thermoregulation to buffer “cold-blooded” animals against climate warming. *Proceedings of the National Academy of Sciences of the United States of America* 106: 3835–3840. Available: <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=2656166&tool=pmcentrez&rendertype=abstract>.
  51. Hoset KS, Espmark Y, Moksnes A, Haugan T, Ingebrigtsen M, et al. (2004) Effect of ambient temperature on food provisioning and reproductive success in snow buntings *Plectrophenax nivalis* in the high arctic. *Ardea* 92: 239–246.
  52. Low M, Eggers S, Arlt D, Part T (2008) Daily patterns of nest visits are correlated with ambient temperature in the Northern Wheatear. *Journal of Ornithology* 149: 515–519. Available: <http://www.springerlink.com/index/10.1007/s10336-008-0300-3>. Accessed 2013 March 21.
  53. Orians GH, Pearson NE (1979) On the theory of central place foraging. *Analysis of Ecological Systems*. Columbus: Ohio State University Press. pp. 155–177.
  54. Sæther B, Andersen R, Pedersen HC (1993) Regulation of parental effort in a long-lived seabird: of the cost of reproduction an experimental manipulation in the Antarctic petrel, *Thalassoica antarctica*. *Behavioral Ecology and Sociobiology* 33: 147–150.
  55. Weimerskirch H, Chastel O, Ackermann L (1995) Adjustment of parental effort to manipulated foraging ability in a pelagic seabird, the thinbilled prion *Pachyptila becheri*. *Behavioral Ecology and Sociobiology* 36: 11–16.
  56. Yahav S, McMurtry JP (2001) Thermotolerance acquisition in broiler chickens by temperature conditioning early in life—the effect of timing and ambient temperature. *Poultry Science* 80: 1662–1666. Available: <http://www.ncbi.nlm.nih.gov/pubmed/11771878>.
  57. Metcalfe NB, Monaghan P (2001) Compensation for a bad start: grow now, pay later? *Trends in Ecology & Evolution* 16: 254–260. Available: <http://www.ncbi.nlm.nih.gov/pubmed/11301155>.
  58. Soobramoney S, Downs CT, Adams NJ (2003) Physiological variability in the fiscal shrike *Lanius collaris* along an altitudinal gradient in South Africa. *Journal of Thermal Biology* 28: 581–594. Available: <http://linkinghub.elsevier.com/retrieve/pii/S0306456503000706>. Accessed 2012 Nov 6.
  59. Porter WP, Gates DM (1969) Thermodynamic equilibria of animals with environment. *Ecological Monographs* 39: 227–244.
  60. Walsberg GE (1993) Thermal consequences of diurnal microhabitat selection in a small bird. *Ornis Scandinavica* 24: 174–182.
  61. Bergmann C (1847) Über die Verhältnisse der Wärmeökonomie der Thiere zu ihrer Grösse. *Göttinger Studien* 1: 595–708.
  62. Gardner JL, Heinsohn R, Joseph L (2009) Shifting latitudinal clines in avian body size correlate with global warming in Australian passerines. *Proceedings of the Royal Society B: Biological Sciences* 276: 3845–3852. Available: <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=2817282&tool=pmcentrez&rendertype=abstract>. Accessed 2012 Nov 5.
  63. Yom-Tov Y (2001) Global warming and body mass decline in Israeli passerine birds. *Proceedings of the Royal Society B: Biological Sciences* 268: 947–952. Available: <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=1088692&tool=pmcentrez&rendertype=abstract>. Accessed 2013 Feb 1.
  64. Gardner JL, Peters A, Kearney MR, Joseph L, Heinsohn R (2011) Declining body size: a third universal response to warming? *Trends in Ecology & Evolution* 26: 285–291. Available: <http://www.ncbi.nlm.nih.gov/pubmed/21470708>. Accessed 2012 Nov 2.
  65. Haywood S, Perrins CM (1992) Is clutch size in birds affected by environmental conditions during growth? *Proceedings of the Royal Society B: Biological Sciences* 249: 195–197. Available: <http://www.ncbi.nlm.nih.gov/pubmed/1360680>.
  66. Green DJ, Cockburn A (2001) Post-fledging care, philopatry and recruitment in brown thornbills. *Journal of Animal Ecology* 70: 505–514. Available: <http://doi.wiley.com/10.1046/j.1365-2656.2001.00503.x>.
  67. Wheelwright NT, Tice KA, Freeman-Gallant CR (2003) Postfledging parental care in savannah sparrows: sex, size and survival. *Animal Behaviour* 65: 435–443.
  68. Magrath RD (1991) Nestling weight and juvenile survival in the blackbird, *Turdus merula*. *Journal of Animal Ecology* 60: 335–351.
  69. Thompson CF, Flux JES (1991) Body mass, composition, and survival of nestling and fledgling starlings (*Sturnus vulgaris*) at Belmont, New Zealand. *New Zealand Journal of Ecology* 15: 41–47.

70. Stienen EWM, Brenninkmeijer A (2002) Variation in growth in sandwich tern chicks *Sterna sandwicensis* and the consequences for pre- and post-fledging mortality. *Ibis* 144: 567–576. Available: <http://doi.wiley.com/10.1046/j.1474-919X.2002.00086.x>.
71. Immelmann K (1973) Role of the environment in reproduction as a source of “predictive” information. In: Farner D, editor. *Breeding Biology of Birds*. Washington DC: National Academy of Sciences. pp. 121–147.